

NUTRIENT CONCENTRATIONS IN BIG CREEK CORRELATE TO REGIONAL WATERSHED LAND USE

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Summary

1. Nutrient concentrations in streams draining the Boston and Ozark Mountains region were related to the intensity of watershed land use.
2. Concentrations in Big Creek were similar to other watersheds in the ecoregion with similar land use, suggesting limited impact of the CAFO on Big Creek at the present time. However, this does not preclude future impacts and longer-term monitoring continues.

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Background

Nutrient impairment of surface waters continues despite widespread conservation efforts to reduce losses from urban, rural, and agricultural land uses (Scavia et al., 2014). Land use within watersheds influences the quality and quantity of water in streams draining the landscape. As land disturbance increases and use intensifies, an increase in stormwater runoff and nutrient inputs that lead to a greater potential for transport to receiving water is generally observed (Dubrovsky et al., 2010; Rebich et al., 2011). This has led to efforts to identify and quantify nutrient sources within watersheds, strategically target, and apportion nutrient loss reduction (Reckhow et al., 2011).

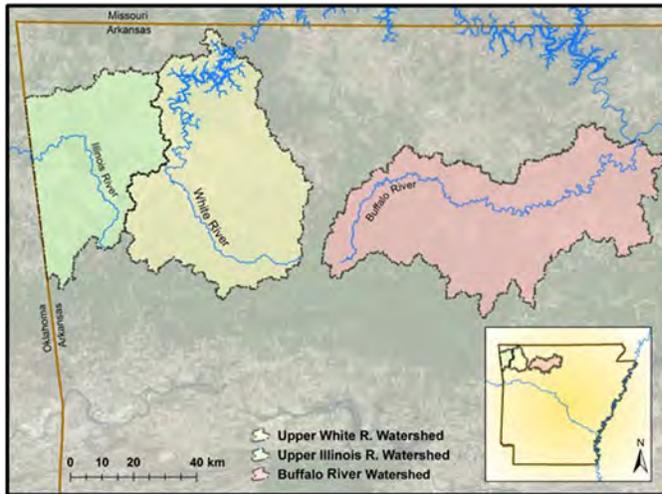
Numerous factors influence the relationship between land use in a given watershed and nutrient transport downstream from that watershed. With an increase in percent of the drainage area in pasture, row crop, and/or urban use, a general trend of increasing nutrient concentrations in storm and base flows will be manifested (Buck et al., 2004; Giovannetti et al., 2013; Haggard et al., 2003; Migliaccio et al., 2007). Thus, nutrient concentrations in streams draining forested lands tend to be less than in watersheds with considerable anthropogenic land use.

Methods

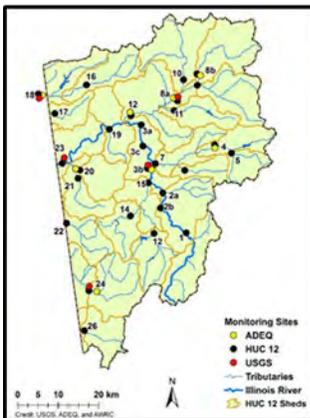
Water samples have been collected over varying periods at the outlet of subwatersheds of the BRW, Upper Illinois River Watershed (UIRW) and Upper White River Watershed (UWRW; Figure 1). Land use and cover (i.e., forest, pasture, and urban) for each subwatershed was obtained from high-resolution (4 m) imagery from U.S. Geological Survey (USGS) National Elevation Dataset (see <https://lta.cr.usgs.gov/NED>; Gesch et al., 2002), National Land-Cover Dataset (see <https://catalog.data.gov/dataset/usgs-national-land-cover-dataset-nlcd-downloadable-data-collection>), and National Hydrologic Dataset (see <https://nhd.usgs.gov/>). In the UWRW, Giovannetti et al. (2013) monitored 20 sites monthly for one year (June 2005 to July 2006), collecting water samples during base-flow conditions. In the UIRW, Haggard et al. (2010) monitored 29 sites monthly during calendar year 2009, also collecting water samples during base-flow conditions.

In the BRW, the National Park Service (NPS) in partnership with the Arkansas Department of Environmental Quality (ADEQ) periodically collected water samples and measured nutrient concentrations at 20 stream sites from 1985 through 2015. Dissolved P, total P, nitrate-N, and total N concentrations were obtained directly from these data. Forest, pasture, and urban land-use areas were determined from 2006 high-resolution (4 m) land use-land cover imagery (Table 1).

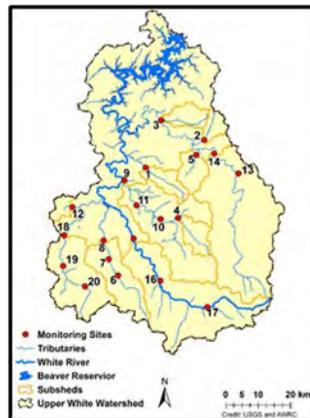
Study watersheds in the Ozark Highlands Ecoregion



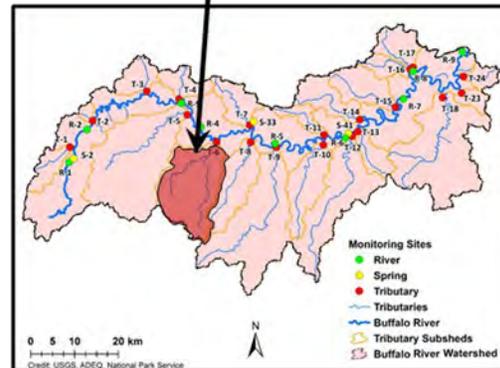
Big Creek Watershed



Upper Illinois River Watershed



Upper White River Watershed



Buffalo River Watershed

Figure 1. Location of the Big Creek, Buffalo River, Upper Illinois River and Upper White River Watersheds in the Boston Mountains and Ozark Highlands ecoregion. Information from U.S. Geological Survey (USGS), Environmental Systems Research Institute (ESRI), and National Aeronautics and Space Administration (NASA).

Big Creek water samples were analyzed at an Arkansas Department of Environmental Quality certified water quality laboratory within the Arkansas Water Resources Center (<http://arkansas-water-center.uark.edu/water-quality-lab.php>), according to methods detailed in Table 2.

The geometric mean of nutrient concentrations of base-flow samples collected between September 2013 and April 2017 were determined, to compare with base-flow nutrient concentrations available for BRW, UIRW, and UWRW. Base-flow conditions in Big Creek were classified from hydrograph inspection when flow had not increased or decreased within three days of sample collection. McCarty and Haggard (2016) suggested that stream nutrient concentrations under base flow can be used to identify nonpoint sources and target remedial measures in Boston Mountains and Ozark Highland watersheds.

Table 1. Percent of forest pasture, and urban land use in the Big Creek, Buffalo River, Upper Illinois, and Upper White Watersheds.

Watershed	Forest	Pasture	Urban
	----- % -----		
Big Creek †			
Upstream	89.5	8.0	2.6
Downstream	79.5	17.0	3.5
Buffalo River	52 - 99	0 - 25	0 - 1
Upper White River	34 - 90	7 - 55	0 – 44
Upper Illinois River	2 - 70	27 – 69	3 – 61

† Up and downstream of CAFO operation and fields permitted to receive manure.

Table 2. Minimum detection limits for each chemical and biological constituent.

Constituent	Analytical method †	Minimum detection limit ‡	Reporting limit ¶
Dissolved P, mg/L	EPA 365.2	0.002	0.010
Total P, mg/L	APHA 4500-P J	0.012	0.020
Nitrate-N, mg/L	EPA 300.0	0.004	0.050
Total N, mg/L	APHA 4500-P J; EPA 365.2	0.006	0.050

† EPA is Environmental Protection Agency Approved CWA Chemical Test Methods, available at <https://www.epa.gov/cwa-methods/approved-cwa-chemical-test-methods#number> and APHA is American Public Health Association from the Wadeable Streams Assessment, Water Chemistry Laboratory Manual http://www.epa.gov/owow/monitoring/wsa/WRS_lab_manual.pdf.

‡ The Minimum detection limit of an analyte is the value, which can be measured and reported with 99% confidence that the analyte concentration is greater than zero. Further information is available at http://water.usgs.gov/owq/OFR_99-193/detection.html

¶ The Reporting limit is the least (non-zero) calibrated standard used in analysis, or as defined by method for total suspended solids.

Using all above-listed data sources, the geometric means of nutrient concentrations for streams in the BRW, UIRW, and UWRW were used to develop a relationship with human development within the

watershed. Human development is defined as the percent of pasture plus urban land use within the watershed. Exponential relationships with 95% confidence bands around the observations were developed for dissolved P, total P, nitrate-N, and total N concentrations to put nutrient concentration at Big Creek into the context of regional stream nutrients and watershed land use.

Data from Big Creek were paired with discharge available from a gaging station just downstream from the swine CAFO, where the USGS developed the rating curve; discharge information was only available from May 2014 through April 2017. The data were then used in a simple three-step process (White et al., 2004) to look at monotonic changes in the nutrients at Big Creek: (1) log-transform concentration (mg/L) and associated instantaneous discharge (m^3/s); (2) use locally weighted regression (LOESS) to smooth the data with a sampling proportion (n) of 0.5; and (3) plot the residuals from LOESS (i.e., the flow-adjusted concentrations), over time and use linear regression to evaluate monotonic trends.

Putting Stream Nutrient Concentrations into Context at Big Creek

In Big Creek, upstream of the swine CAFO, geometric mean concentrations of base flow for the monitoring period extending from September 13, 2013 to July 11, 2019 (121 samples) for dissolved P, total P, nitrate-N, and total N were 0.008, 0.025, 0.10, and 0.19 mg/L, respectively. Directly downstream of the CAFO, geometric mean concentrations at Big Creek during base flow conditions during the same period (151 samples) were 0.010, 0.024, 0.27, and 0.38, mg/L for dissolved P, total P, nitrate-N, and total N, respectively. Arkansas has narrative criteria for nutrient concentrations in streams (Arkansas Pollution Control and Ecology Commission, 2016), but its proposed assessment methodology has numeric screening concentrations for total N (0.45 - 2.43 mg/L) and total P (0.04 - 0.10 mg/L) in the Boston Mountains and Ozark Highlands. The geometric mean concentrations at Big Creek upstream and downstream from the CAFO were below these values for the Boston Mountains and Ozark Highlands ecoregion.

Nutrient concentrations in Big Creek upstream and downstream from the CAFO are low with respect to nutrient–biological–response thresholds for algae, macroinvertebrates and fish. Evan-White et al. (2014) reviewed the literature, summarizing nutrient–biological–response thresholds across the U.S.:

- *Algal Metric Responses* total N: 0.38–1.79 mg/L total P: 0.011–0.28 mg/L
- *Macroinvertebrate Metric Responses* total N: 0.61–1.92 mg/L P: 0.04–0.15 mg/L
- *Fish Metric Responses* total N: 0.54–1.83 mg/L total P: 0.06–0.14 mg/L

Total N concentrations at Big Creek upstream and downstream of the swine CAFO were well below thresholds that result in some expected biological response, whereas total P concentrations were below thresholds for expected macroinvertebrate and fish response and on the low end of the range for expected algal response. However, these lower total P thresholds (0.006 - 0.026 mg/L; Stevenson et al., 2008) were focused on shifts in diatom species and metrics rather than nuisance algal biomass. A recent study on the Illinois River Watershed showed that stream total P thresholds with *Cladophora* biovolume and nuisance taxa proportion of biovolume were observed between 0.032 and 0.058 mg/L (Joint Study

Committee, 2017). Thus, total P concentrations at Big Creek upstream and downstream of the CAFO were in the range where the natural assemblage of algae is shifting, but these concentrations would likely not be indicative of problematic nuisance algae in this ecoregion.

Geometric mean nutrient concentrations varied upstream and downstream of the swine CAFO at Big Creek, and Kusic et al. (2015) used the publicly available data to allude to the N increase being from human activities on the landscape, e.g., the CAFO. However, the historic land use and how stream nutrient concentrations during base-flow conditions increase with human development within the Boston Mountain and Ozark Highland watersheds need to be considered (e.g., see Giovannetti et al., 2013; Haggard et al., 2003; Migliaccio et al., 2007). In the Big Creek watershed, the percent of land influenced by human activities (i.e., pasture plus urban) doubles from ~10 to ~20% in the drainage area upstream and downstream of the CAFO. Nutrient concentrations in Big Creek upstream and downstream of the CAFO are within the range typical of streams draining similar land uses (Figure 2).

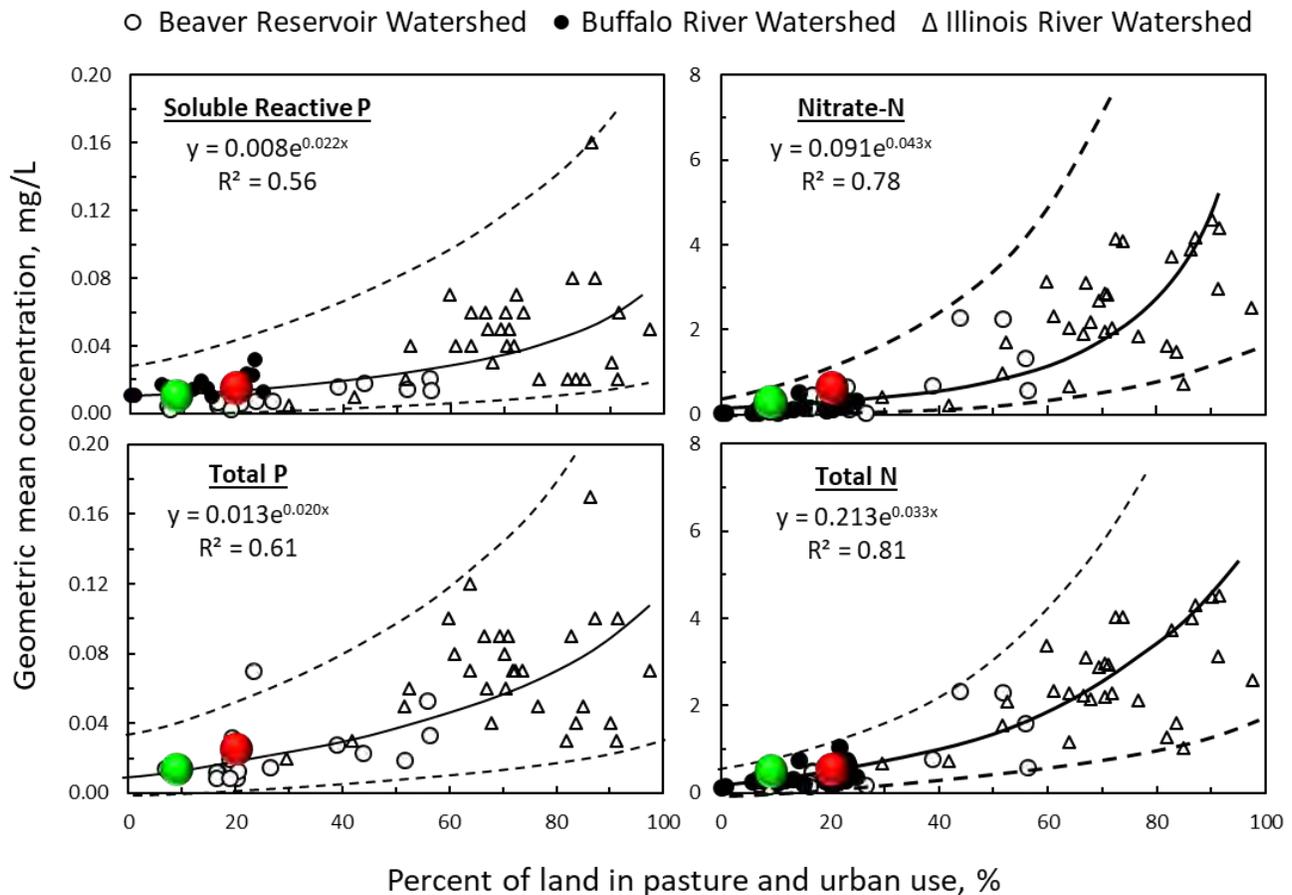


Figure 2. Relationship between land use and the geometric mean N and P concentrations (mg/L) in the Buffalo, Upper Illinois, and Upper White River Watersheds. Dashed lines represent the 95% confidence intervals for the estimated mean (solid line). Green points are geometric mean concentration measured upstream of the CAFO on Big Creek and red points are geometric mean concentration measured downstream of the CAFO on Big Creek.

At this time, nutrient concentrations in Big Creek upstream and downstream from the swine CAFO are consistent with the range in concentrations for other watersheds with similar pasture and urban land use characteristics (Figure 2), as well as less than most nutrient thresholds for nuisance water-quality conditions (Omernik and Griffith, 2014). However, this does not preclude the possibility that nutrient concentrations at Big Creek may increase over time especially if human development and activity in the drainage areas increase. The most important observation is that nutrient concentrations were low in Big Creek providing the ability to detect changes over time.

First, understanding that long-term (e.g., decadal scale) water-quality data are needed to reliably assess how stream nutrient concentrations have changed in response to watershed management and climate variations is of critical importance (Hirsch et al., 2015). The literature shows that stream nutrient concentrations can change relatively quickly in response to effluent management (e.g., Haggard, 2010; Scott et al., 2011), but seeing a response (i.e., decrease in concentrations) from landscape management can take decades or more (Green et al., 2015; Sharpley et al., 2013). A myriad of factors may influence observed nutrient concentrations in streams, including discharge (Petersen et al., 1998), biological processes and climatic conditions (i.e., drought and floods; Jones and Stanley, 2016), and dominant transport pathways (Sharpley et al., 2013). Thus, we need to use caution when interpreting trends in water quality over databases that only cover a limited timeframe.

Nutrient concentrations at Big Creek upstream and downstream of the swine CAFO, and indeed most tributaries of the Buffalo River, are low relative to other watersheds in this ecoregion (Figure 2). This provides a starting point to build a framework to evaluate changes in nutrient concentrations of streams as a function of land use and management and to establish baseline, in-stream nutrient concentrations and a process by which time and/or land use and management impacts can be determined.

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